

Calcium-carboxyl intra-bridging during interfacial polymerization: A novel strategy to improve antifouling performance of thin film composite membranes

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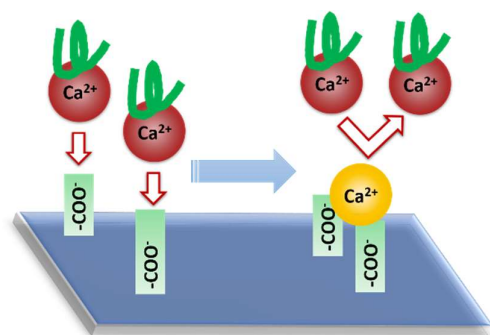
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27 **ABSTRACT**

28 This study reports a novel intra-bridging strategy to improve the antifouling
29 performance of a thin-film composite (TFC) membrane. We demonstrate that the
30 addition of Ca^{2+} during the interfacial polymerization reaction led to the formation of
31 stable Ca^{2+} -carboxyl complexes within the polyamide rejection layer. This intra-
32 bridging of carboxyl groups by Ca^{2+} effectively sequestered them, reducing their
33 availability for binding divalent metal ions in the aqueous solution and for forming
34 foulant-metal-membrane inter-bridges. Membrane fouling and cleaning experiments
35 confirmed improved flux stability and fouling reversibility for the Ca^{2+} modified
36 membranes. The greatly enhanced antifouling performance of these membranes,
37 together with their better surface hydrophilicity and greater water permeability, makes
38 the intra-bridging approach highly attractive in overcoming the classical permeability-
39 selectivity-antifouling tradeoff. Our findings pave a new direction for synthesizing
40 high-performance TFC membranes.

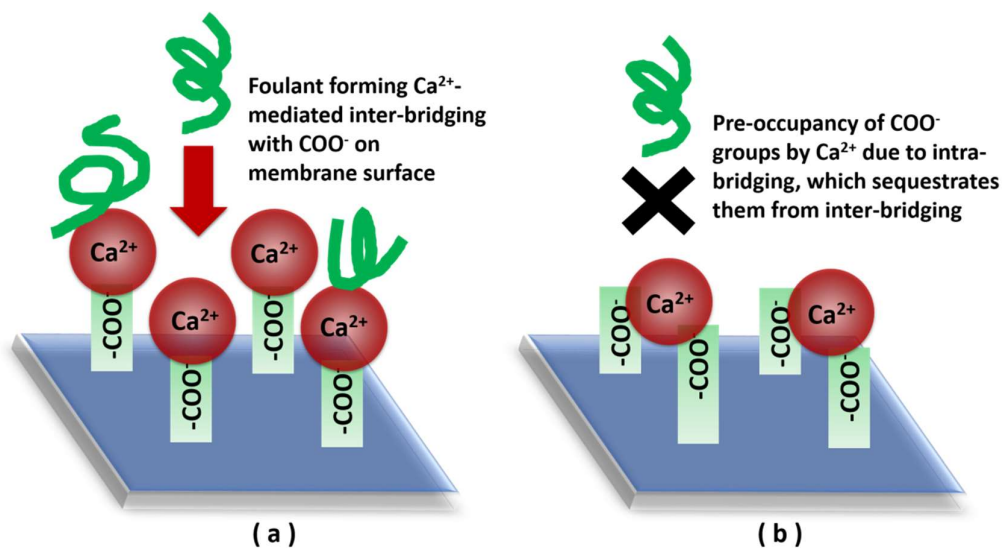
41 INTRODUCTION

42 Reverse osmosis membranes applied for desalination and water reuse are dominated by
43 thin film composite (TFC) polyamide chemistry.¹⁻³ These membranes comprise a thin
44 polyamide rejection layer on a porous support layer, in which the polyamide rejection
45 layer determines not only the membrane separation properties but also their fouling
46 propensity.^{2, 4} Although the state-of-the-art TFC membranes possess desirable
47 combinations of salt rejection and water permeability, their polyamide chemistry is
48 often prone to membrane fouling.¹⁻⁶

49 One of the key factors contributing to TFC membranes' vulnerability to fouling is the
50 extensive presence of carboxyl groups on their surface.^{4, 7-14} During the interfacial
51 polymerization (IP) reaction of amine and acyl chloride monomers for the fabrication
52 of TFC membrane (e.g., 1,3-phenylenediamine (MPD) and 1,3,5-benzenetricarbonyl
53 trichloride (TMC), respectively), the unreacted acyl chlorides in the polyamide matrix
54 rapidly hydrolyze into carboxyl groups.^{7, 8, 14-16} These carboxyl groups are susceptible
55 to organic fouling in the presence of divalent ions such as Ca^{2+} (Figure 1a).¹⁷⁻²⁴ Ca^{2+}
56 ions bind to the carboxyl groups on the membrane surface and those of the organic
57 foulants to form membrane- Ca^{2+} -organic bridges (referred as the inter-bridges in the
58 current study), which often initiates severe foulant accumulation on the membrane
59 surface.^{17, 25} The presence of carboxyl groups also promotes biofouling^{26, 27} and
60 scaling^{28, 29}. Although surface modifications methods such as coating or grafting an
61 antifouling surface layer³⁰⁻³² can address this issue, it is often at the expense of reduced
62 membrane permeability, higher cost and/or complex procedures.

63 For the first time, we propose a novel strategy to address the poor fouling performance
64 of TFC membranes by sequestering their carboxyl groups (Figure 1b). We hypothesize
65 that the ability of carboxyl groups for promoting membrane- Ca^{2+} -organic inter-bridges

66 can be significantly weakened by “pre-occupancy” of these vulnerable sites during the
 67 membrane formation stage. Specifically, we demonstrate this idea with the simple
 68 addition of Ca^{2+} into the MPD monomer solution before performing the IP reaction.
 69 The presence of Ca^{2+} during the IP reaction promotes the formation of carboxyl- Ca^{2+} -
 70 carboxyl bridges within the polyamide layer. This intra-bridging by Ca^{2+} pre-occupies
 71 the carboxyl groups in the polyamide layer, making them unavailable to the formation
 72 of inter-bridges with foulants and thus minimizing membrane fouling propensity.
 73 In the current study, we prepared a Ca^{2+} -intra-bridged TFC membrane and
 74 systematically characterized its surface composition, separation properties, and
 75 antifouling performance. We demonstrate superior antifouling performance of this
 76 novel membrane under reverse osmosis conditions. The novel intra-bridging strategy
 77 reported in this work provides an easy and efficient pathway for enhancing antifouling
 78 performance of TFC membranes.



79
 80 **Figure 1. Illustration of anti-fouling mechanism of modified TFC membrane by Ca^{2+} addition**
 81 **during IP process. (a) TFC-control membrane. (b) Ca^{2+} -modified TFC membrane.**

82

83 MATERIALS AND METHODS

84 **Materials and Chemicals.** Unless otherwise specified, all chemicals and reagents used
85 in this study were of analytical grade. Udel[®] polysulfone (PSf, Mn=143 kDa, η =1.01
86 dL g⁻¹) was obtained from Solvay as the polymer material. Solvent N-methyl-2-
87 pyrrolidone (NMP, Sigma-Aldrich, >99.5%) and pore former diethylene glycol (DEG,
88 Sigma-Aldrich, >99.0%) were used to fabricate the PSf substrate. MPD (Sigma-
89 Aldrich, >99%) and TMC (Sigma-Aldrich, >98%) were employed to synthesize the
90 polyamide active layer via interfacial polymerization. Sodium chloride (NaCl), calcium
91 chlorides (CaCl₂), magnesium chloride (MgCl₂) and barium chloride (BaCl₂) were
92 obtained from Tianjin Kemiou Chemical Reagent Co., Ltd., China. Ethylene glycol,
93 diethylene glycol, and glucose were obtained from Tianjin Kemiou Chemical Reagent
94 Co., Ltd., China. Milli-Q water with a resistivity of 18.2 MΩ cm was used to prepare
95 all working solutions. For membrane fouling evaluation, sodium alginate (SA, Sigma-
96 Aldrich) and tannic acid (TA, Sigma-Aldrich) were selected as the model organic
97 foulants to represent polysaccharides and humic substances, respectively. The stock
98 solutions for the foulants were prepared by dissolving the foulants' power in Milli-Q
99 water. All stock solutions were kept in the dark at 4 °C and used within one month after
100 their preparation.

101
102 **Preparation of Ca²⁺-inter-bridged TFC Membrane.** To prepare the PSf substrate, the
103 polymer (PSf, 15 wt.%), pore former (DEG, 17 wt.%) and solvent (NMP, 68 wt.%)
104 were stirred for 12 h for complete dissolution. The polymer solution was then degassed
105 for more than 24 h. The resultant homogeneous solution was spread on a glass plate
106 with a casting knife of 100 μm in height. The nascent film was immediately immersed
107 into a water bath at room temperature (23 °C) for 10 min. The resultant PSf substrate
108 was washed thoroughly and stored in DI water before use.

109 Interfacial polymerization was performed on the PSf substrate.³³ The PSf substrate was
110 first immersed in 3.4 wt.% MPD aqueous solution for 120 s. To allow Ca^{2+} -intra-
111 bridging, different amount of CaCl_2 (0, 0.2, 0.5, 1.0 and 2.0 wt.%) were dosed in the
112 MPD solution. After the MPD soaking step, the excess MPD solution was removed
113 using an air knife. A 0.15 wt.% TMC n-hexane solution was then brought into contact
114 with the MPD saturated substrate for 60 s to form the polyamide layer. The nascent
115 membrane was cured in DI water at 90 °C for 120 s, then rinsed with a 200 ppm NaClO
116 aqueous solution for 120 s, followed by rinsing with a 1000 ppm NaHSO_3 aqueous
117 solution for 30 s. Finally, the prepared Ca^{2+} -modified TFC membranes were cured at
118 90 °C for 120 s and stored in DI water at 4 °C before testing. The resulting membranes
119 were denoted as TFC- Ca^{2+} (W), where W is the concentration of CaCl_2 (in wt.%) dosed
120 in the MPD solution (Table S1, Supporting Information S1).

121

122 **Characterization of the Ca^{2+} -modified Membranes.** The surface functional groups
123 of the Ca^{2+} -modified membranes were characterized through Attenuated Total Internal
124 Reflectance Fourier Transform Infrared Spectrometer (ATR-FTIR, Thermo
125 ESCALAB250, USA). The chemical element composition of the Ca^{2+} -modified
126 polyamide layer was analyzed using X-ray photoelectron spectroscopy (XPS, Axis
127 Ultra DLD, Kratos X-ray Photoelectron Spectrometer). The surface hydrophilicity of
128 the fabricated membrane was evaluated by a contact angle meter (SL200B, Solon Tech
129 Co., Ltd., China). The zeta potential of the TFC membranes was measured by an electro
130 kinetic analyzer (Anton Paar GmbH, Austria) with the streaming potential
131 measurements. An atomic force microscope (AFM, BioScope, Veeco Inc.) was used to
132 characterize the membrane surface morphology.

133 The carboxyl group density was measured by the toluidine blue (TBO) method

134 developed by Elimelech and co-workers.^{7, 8} Briefly, membrane coupons of 2 cm² were
135 contacted with 2 mM TBO solution at pH 11 for 3 min. After rinsing with and
136 immersion in a dye-free NaOH solution (pH = 11) for more than 4 h, the membrane
137 coupons were placed into 10 mL of 0.2 M NaCl solutions at pH 2 for 30 min under
138 stirring to release the bound TBO into the solution. The resultant solutions were
139 analyzed by optical density at a 630 nm wavelength. Light absorption was converted to
140 TBO concentration to calculate the membrane surface carboxyl group density.^{7, 8}

141

142 **Performance Evaluation of the Ca²⁺-modified TFC Membranes.**

143 **Separation performance characterization.** The water permeability (A), salt
144 permeability (B), and salt rejection (R) of the TFC membranes were measured in a
145 crossflow reverse osmosis filtration setup at room temperature (23 ± 0.5 °C). Briefly, a
146 membrane (effective area of 15.89 cm²) was first compacted with DI water at a cross
147 flow velocity of 8.5 cm/s ($Re \approx 174$) and an applied pressure of 20 bar until the permeate
148 flux became stable. The intrinsic water permeability coefficient, A , was then calculated
149 by dividing the measured water flux J_w by the applied pressure ΔP , i.e., $A = J_w/\Delta P$.
150 Subsequently, a 20 mM NaCl was introduced to the feed solution to determine the salt
151 rejection. The observed NaCl rejection, R , was calculated from $R = 1 - C_p/C_f$, where the
152 permeate concentration C_p and the feed concentration C_f were both based on
153 conductivity measurements. Neutral hydrophilic molecules, including ethylene glycol,
154 diethylene glycol, and glucose (MW = 62.1, 106.1, and 180.2 Dalton, respectively),
155 were also used to evaluate the effect of membrane modification on the size exclusion
156 effect.³⁴⁻³⁶ Rejection of each compound was determined using a feed concentration of
157 200 mg/L under a pressure of 20 bar, and both the feed and permeate concentration
158 were determined by total organic carbon tests.

159 **Ca²⁺ adsorption tests.** To verify the effect of intra-bridging on the sequestration of
160 carboxyl groups, we performed Ca²⁺ adsorption tests for the TFC-control and TFC-Ca²⁺
161 (1.0) membranes. Membrane coupons with an area of $1.0 \times 2.0 \text{ cm}^2$ were soaked in 15
162 mL 1 wt.% CaCl₂ solutions (pH adjusted to 3, 4, 5, and 7) at 4 °C. The water samples
163 were collected at 1, 3, 5, and 7 days, and the amount of Ca²⁺ adsorbed on TFC-control
164 and TFC-Ca²⁺ (1.0) membranes was quantified by an inductively coupled plasma
165 optical emission spectrometer (ICP-OES, 5300DV, PerkinElmer). The average values
166 from three membrane samples were reported.

167 **Membrane fouling characterization.**

168 Membrane fouling and cleaning were performed using sodium alginate (SA) and tannic
169 acid (TA) as model organic foulants to represent polysaccharides and humic substances,
170 respectively. Briefly, a new membrane was first equilibrated with the foulant-free
171 background electrolyte (1 mM CaCl₂ + 47 mM NaCl) under pressure until a stable water
172 flux achieved. To ensure that fouling behavior can be compared between different
173 membranes, the applied pressure was adjusted to obtain an identical initial water flux
174 of 22 LMH for all fouling experiments. To start membrane fouling, a foulant stock
175 solution was added to the feed tank to reach a final feed foulant concentration of 200
176 mg/L of SA or TA. During the entire test, the following conditions were maintained:
177 feed solution pH 7.2 ± 0.2 , ionic strength of 50 mM (1 mM CaCl₂ + 47 mM NaCl),
178 constant temperature of $23 \pm 0.5 \text{ °C}$, and a cross flow velocity of 8.5 cm/s ($Re \approx 174$).
179 Physical cleaning with DI water was also performed at the end of the fouling
180 experiments to evaluate fouling reversibility.³⁷⁻³⁹ Membrane cleaning involved DI
181 water rinsing at the same cross-flow velocity of 8.5 cm/s for 60 min. After cleaning,
182 pure water flux of the tested membrane was measured to calculate the flux recovery
183 ratio. Each fouling and cleaning experiment was repeated at least twice using different

184 membrane coupons.

185

186 **RESULTS AND DISCUSSION**

187 **Membrane Characterization.** Figure 2 presents ATR-FTIR, XPS, contact angle and
188 zeta potential results. The ATR-FTIR spectra (Figure 2a) of both the TFC-control and
189 the Ca^{2+} -modified membranes show the characteristic peaks of fully aromatic
190 polyamide: 1540 cm^{-1} (amide II peak, assigned to the bending vibration of $-\text{N}-\text{H}$),
191 1660 cm^{-1} (amide I peak, assigned to the stretching vibration of $-\text{C}=\text{O}$), and 1613 cm^{-1}
192 (ascribed to the bending vibration of aromatic ring in the polyamide active layer).^{16, 37}
193 However, a new peak occurred at 1728 cm^{-1} (carbonyl group in stretching mode ⁴⁰) for
194 the Ca^{2+} -modified membranes, which can be attributed to the complexation of Ca^{2+} with
195 the carboxyl groups on the membrane surface. Consistent with the FTIR results, XPS
196 results (Figure 2b and Table S2) confirmed increased presence of calcium (Ca2p peak
197 at the binding energy of 348 eV) in the polyamide rejection layer in the order of TFC-
198 control (0.0 %) < TFC- Ca^{2+} (0.2) (0.1 %) < TFC- Ca^{2+} (0.5) (0.3 %) < TFC- Ca^{2+} (1.0)
199 (0.5 %). However, TFC- Ca^{2+} (2.0) had a lower Ca^{2+} content (0.3%) compared to TFC-
200 Ca^{2+} (1.0), indicating that further increasing the CaCl_2 concentration was not effective
201 for incorporating more Ca^{2+} in the polyamide layer.

202 Water contact angle (CA) measurements (Figure 2c) show improved surface
203 hydrophilicity upon Ca^{2+} modification. The TFC-control membrane presents the most
204 hydrophobic membrane surface with the water contact angle of 73.9° , which is in line
205 with previously published data for a hand-cast polyamide membrane.¹⁹ The water
206 contact angle dropped sharply to 29.8° for TFC- Ca^{2+} (1.0), which is accompanied with
207 a noticeably reduction in the surface roughness and a shift from leaf-like to nodular
208 surface morphology (Figure 3). Although the underlining mechanism is not clearly

known, the dramatic changes in surface hydrophilicity and morphology may be ascribed to a significant change of the physiochemical environment within the Ca^{2+} -modified membranes (e.g., changed hydrogen bonding behavior and the ability of Ca^{2+} to accommodate hydration water molecules). In addition, our recent study revealed that the surface roughness of TFC polyamide membranes is controlled by the degassing of nanosized CO_2 gas bubbles from the aqueous amine solution.^{41, 42} The formation of Ca^{2+} -carboxyl complexation in the current study may significantly change this degassing behavior and thus the surface morphology. Zeta potential (ζ) results show a less negative membrane surface with the increase of Ca^{2+} concentration in MPD aqueous solution up to 1 wt% (-19.4 ± 0.9 mV for the TFC- Ca^{2+} (1.0) membrane vs. -28.6 ± 1.2 mV for the TFC-control membrane), which is attributed to the partial charge neutralization effect upon Ca^{2+} -carboxyl complexation.⁴³ To determine the density of the free carboxyl groups, the TBO method^{7, 8} was used (Figure 2e). The TFC- Ca^{2+} (1.0) membrane had a carboxyl density of $16.9 \pm 0.5 \text{ nm}^{-2}$, or approximately half of that of the TFC-control membrane. These results not only confirm the effectiveness on the sequestration of carboxyl groups by the Ca^{2+} intra-bridging strategy but also explain the reduced zeta potential for TFC- Ca^{2+} (1.0). Compared to TFC- Ca^{2+} (1.0), TFC- Ca^{2+} (2.0) had a rougher surface (Figure 3e,f), greater water contact angle (Figure 2c), and more negative surface charge (Figure 2d), which is consistent with its lower Ca^{2+} incorporation (Figure 2b and Table S2).

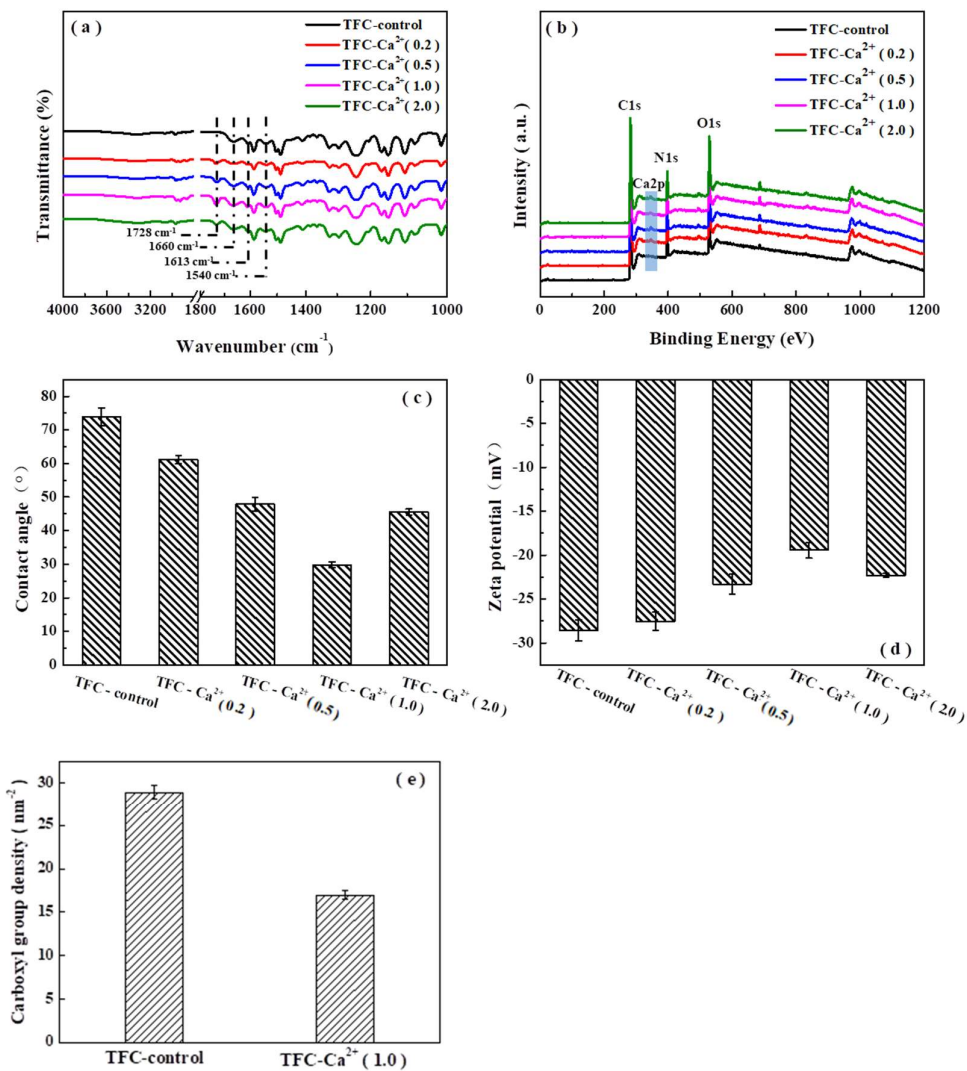


Figure 2. Surface characterization of the TFC-control and Ca^{2+} -modified TFC membranes:
 (a) ATR-FTIR spectra, (b) XPS spectra, (c) water contact angle results, (d) zeta potential results (at pH 7.4), and (e) carboxyl group density.

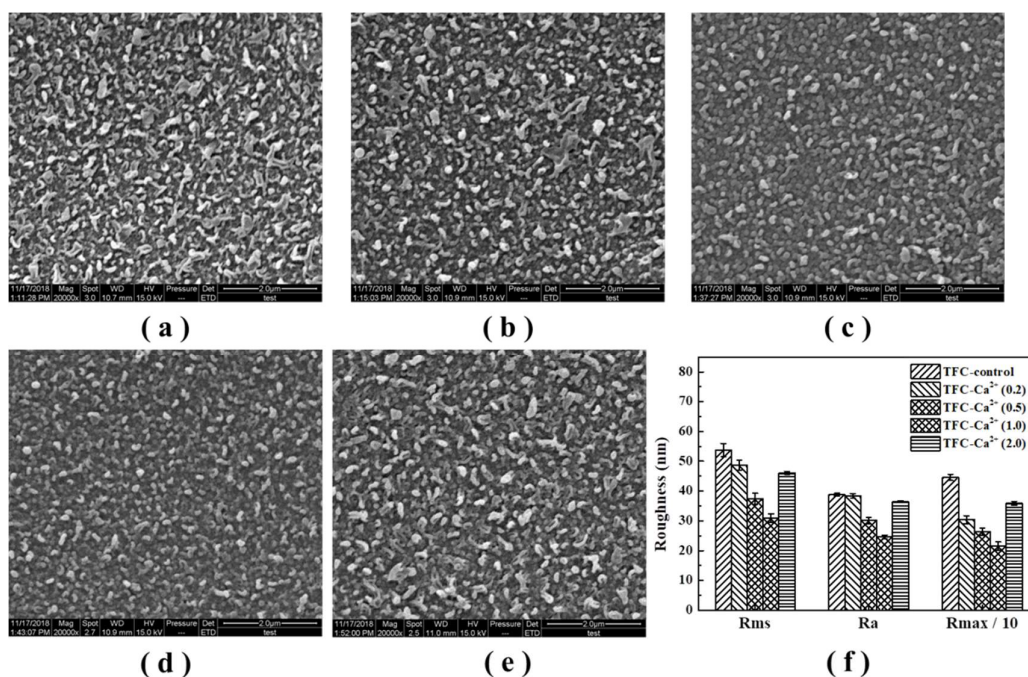


Figure 3. Microscopic characterization. SEM images of the polyamide active layer of (a) TFC-control, (b) TFC-Ca²⁺ (0.2), (c) TFC-Ca²⁺ (0.5), (d) TFC-Ca²⁺ (1.0), (e) TFC-Ca²⁺ (2.0), and (f) their AFM root mean square (Rms), average (Ra) and maximum (Rmax) surface roughness.

Performance of the Ca²⁺-modified TFC Membranes. Table 1 presents the membrane separation properties (i.e., water permeability (*A*), salt permeability (*B*), and salt rejection (*R*)) obtained by reverse osmosis testing. The TFC-control membrane had the lowest *A* value of $1.70 \pm 0.11 \text{ L} \cdot \text{m}^{-2} \cdot \text{h}^{-1} \cdot \text{bar}^{-1}$. In comparison, the TFC-Ca²⁺ (1.0) membrane show a higher water permeability of $2.44 \pm 0.09 \text{ L} \cdot \text{m}^{-2} \cdot \text{h}^{-1} \cdot \text{bar}^{-1}$. The increase in water permeability is ascribed to the significantly improved hydrophilicity of the Ca²⁺-modified polyamide layer (Figure 2c). Our results are consistent with the study by Wan et al.,⁴⁴ who performed interfacial polymerization on a calcium-chloride-dosed polyethersulfone substrate and found that the resulting polyamide became permeable.

The increase in water permeability was accompanied with an increase in salt

251 permeability ($0.42 \text{ Lm}^{-2}\text{h}^{-1}$ for TFC- Ca^{2+} (1.0) vs. $0.37 \text{ Lm}^{-2}\text{h}^{-1}$ for TFC-control, Table
252 1). Nevertheless, all the membranes maintained comparable salt rejection ($\sim 98\%$). Salt
253 rejection of TFC membranes are affected by the combined effects of size exclusion and
254 charge repulsion. To assess the size exclusion effect, additional rejection tests were
255 performed for TFC-control and TFC- Ca^{2+} (1.0) using three neutral hydrophilic
256 molecules (ethylene glycol, diethylene glycol, and glucose, see Figure S1 in Supporting
257 Information S2).³⁴⁻³⁶ Since these molecular probes are neutral, their higher rejections
258 by TFC- Ca^{2+} (1.0) suggest the formation of a tighter (i.e., more crosslinked) polyamide
259 layer with enhanced size exclusion effect, which could be attributed to the formation of
260 carboxyl- Ca^{2+} -carboxyl intra-bridging (i.e., metal-ligand complex) in addition to the
261 covalent amide bonds. While the enhanced size exclusion effect had the tendency to
262 increase salt rejection, the charge neutralization induced by the intra-bridging (Figure
263 2d) would reduce the charge repulsion effect with a negative impact on salt rejection.
264 As a result of these competing effects, the overall NaCl rejection was not significantly
265 affected. Figure 4 compares the separation performance of the Ca^{2+} -modified
266 membranes with commercial reverse osmosis membranes. The separation performance
267 of TFC- Ca^{2+} (1.0) was on par with commercial benchmarks.⁴⁵
268

Table 1 Transport Properties of the TFC-control membrane and the Ca^{2+} -modified membranes.

| Sample | Water permeability (A , $\text{Lm}^{-2}\text{h}^{-1}\text{bar}^{-1}$) | Salt permeability (B , $\text{Lm}^{-2}\text{h}^{-1}$) | Salt rejection (R , %) |
|-------------------------------|--|--|---------------------------|
| TFC-control | 1.70 ± 0.11 | 0.37 ± 0.06 | 98.1 ± 0.5 |
| TFC- Ca^{2+} (0.2) | 1.72 ± 0.10 | 0.39 ± 0.09 | 98.0 ± 0.3 |
| TFC- Ca^{2+} (0.5) | 1.93 ± 0.13 | 0.39 ± 0.10 | 97.7 ± 0.3 |
| TFC- Ca^{2+} (1.0) | 2.44 ± 0.09 | 0.42 ± 0.05 | 97.9 ± 0.3 |
| TFC- Ca^{2+} (2.0) | 1.97 ± 0.15 | 0.40 ± 0.07 | 98.0 ± 0.2 |

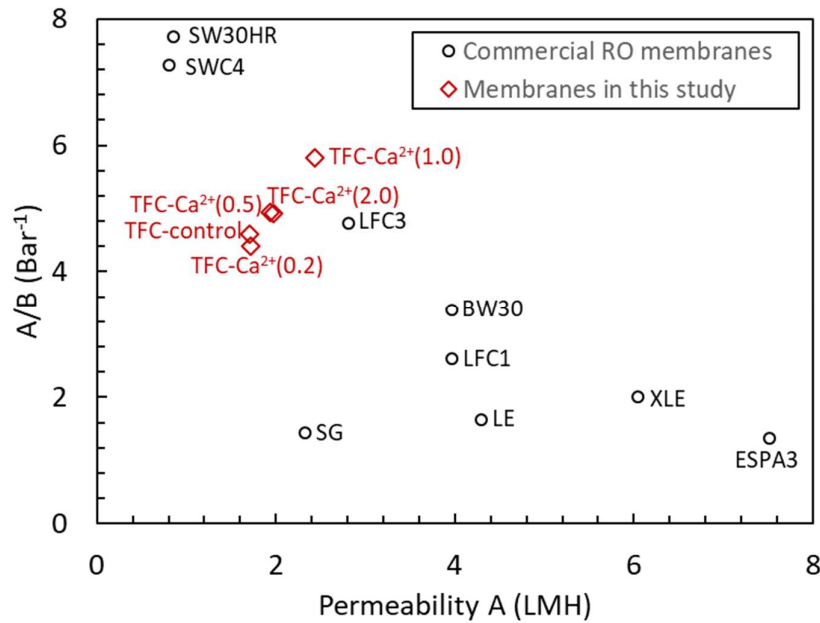


Figure 4. Benchmarking performance of the Ca^{2+} -modified membranes with commercial reverse osmosis membranes. The separation properties of commercial RO membranes were obtained from references ^{45, 46}.

Antifouling Capability of the Ca^{2+} -modified Membranes. Figure 5 presents the fouling behavior of the TFC-control membrane and the TFC- Ca^{2+} (1.0) membrane

279 using a feed solution containing 200 mg/L sodium alginate (SA) or tannic acid (TA).
280 The control membrane experienced significant flux decline (Figure 5a,b). In contrast,
281 the Ca^{2+} -modified TFC membranes showed significantly improved flux stability. The
282 TFC- Ca^{2+} (1.0) membrane exhibits only a mild flux decline of 13.2% for sodium
283 alginate, compared to 24.8% for the TFC-control membrane (Figure 5a). A similar trend
284 was observed when tannic acid was used as a model foulant (Figure 5b). The reduced
285 fouling tendency of the Ca^{2+} -modified membranes can be attributed to the sequestration
286 of the carboxyl groups due to their intra-bridging with Ca^{2+} (Figure 1), in addition to
287 the more hydrophilic and smoother membrane surfaces.

288 Fouling reversibility was evaluated by cleaning the fouled membrane with DI water for
289 60 min. Compared to the control membrane, the TFC- Ca^{2+} (1.0) membrane had more
290 reversible fouling, with 94.4% and 98.2% of the water flux recovered using sodium
291 alginate and tannic acid as model foulants, respectively. Since TFC polyamide
292 membrane is also the dominant membrane type used for osmotically-driven membrane
293 process,⁴⁷⁻⁵¹ we further evaluated the antifouling performance of the calcium modified
294 membranes under forward osmosis testing conditions.^{13, 24, 52} Figure S3 and S4
295 (Supporting Information 3) clearly demonstrate improved flux stability for the TFC-
296 Ca^{2+} (1.0) membrane even under forward osmosis fouling. Once again, this superior
297 fouling reversibility can be explained by of the sequestration of the carboxylic groups
298 for the Ca^{2+} -modified membranes, which weakens the binding between the polyamide
299 layer and the foulants.

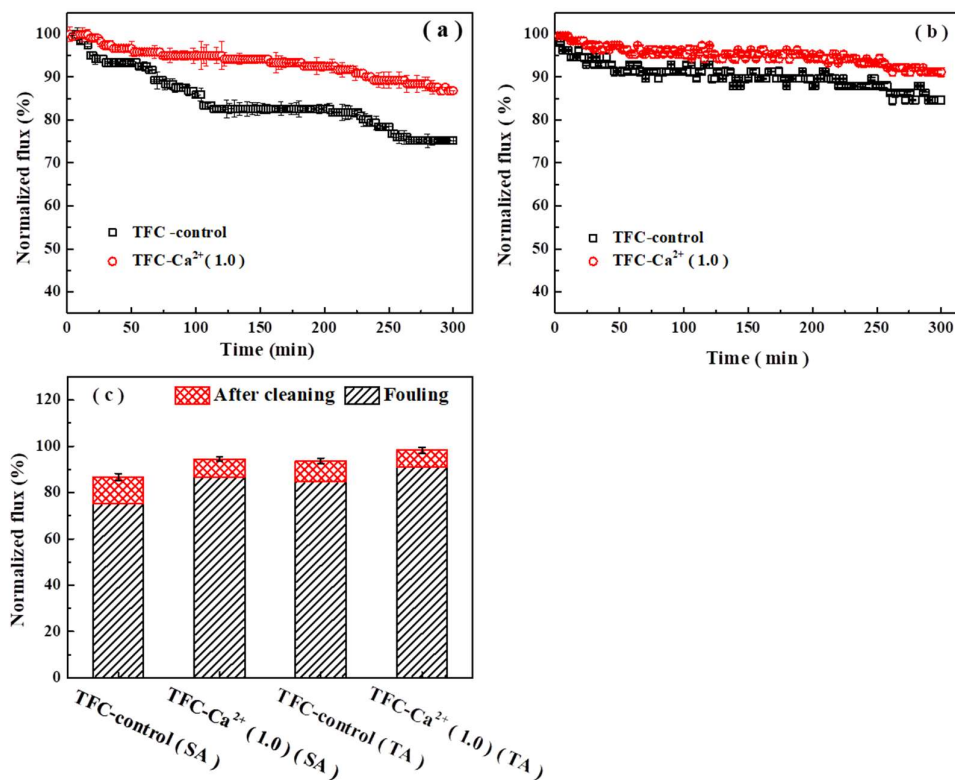
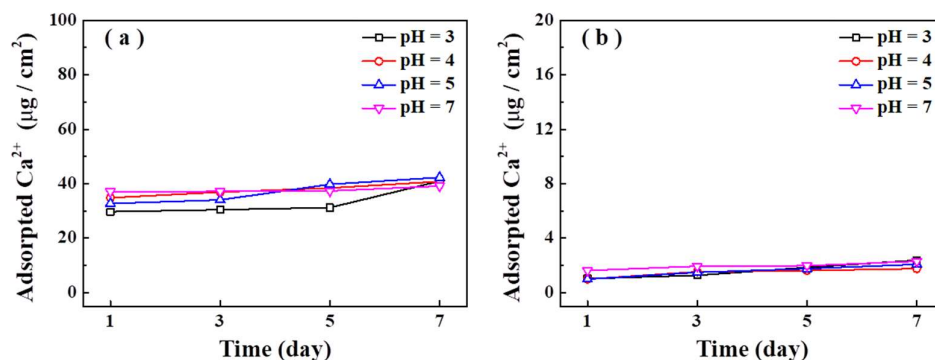


Figure 5. Reverse osmosis membrane fouling and fouling reversibility for the TFC-control and TFC-Ca²⁺ (1.0) membranes using sodium alginate (SA, 200 mg/L) and tannic acid (TA, 200 mg/L) as model foulants. (a) SA fouling tests; (b) TA fouling tests; and (c) fouling reversibility tests. An identical initial water flux of 22 LMH was used for all fouling experiments. Fouling reversibility was evaluated by cleaning the fouling membranes with DI water under crossflow conditions. Error bars are based on the range of the two repeated tests.

To further confirm the sequestration of carboxyl groups by the intra-bridging strategy, we investigated the ability of the TFC-control and TFC-Ca²⁺ (1.0) membranes to adsorb divalent ions (Ca²⁺, Mg²⁺, and Ba²⁺) from the bulk solution. Ca²⁺ uptake (Figure 6) was measured under low to neutral pH (pH 3 - 7), since it is more easily leached (or less

314 tightly bound) under these conditions.^{53, 54} The Ca^{2+} uptake by the TFC-control
 315 membrane was approximately $42.3 \mu\text{g}/\text{cm}^2$, which is an order of magnitude greater than
 316 that by the TFC- Ca^{2+} (1.0) membrane (approximately $3.4 \mu\text{g}/\text{cm}^2$). Similar trends were
 317 observed for the uptake of Mg^{2+} and Ba^{2+} , respectively (Supporting Information S4.1).
 318 In the presence of these divalent ions, the TFC- Ca^{2+} (1.0) membrane once again had
 319 better antifouling performance compared to the TFC-control membrane (Supporting
 320 Information S4.2). These results provide strong evidence that the intra-bridging
 321 approach was highly effective in reducing the availability of the carboxyl groups
 322 contained in the TFC- Ca^{2+} (1.0) membrane, which plays a critical role in enhancing its
 323 antifouling performance (Figure 5).



324
 325 **Figure 6. Uptake of Ca^{2+} by the TFC-control (a) and TFC- Ca^{2+} (1.0) membranes (b) by**
 326 **immersing them in a 1 wt.% CaCl_2 solution.**

327
 328 **Membrane Stability.** To evaluate the stability of the Ca^{2+} -modified membrane, we
 329 performed Ca^{2+} leaching tests for the TFC- Ca^{2+} (1.0) membrane in aqueous solutions at
 330 pH 3, 4, 5 and 7 over a 7-day period (Figure S7 in Supporting Information S5). Even at
 331 the lowest pH of 3, the released Ca^{2+} was significantly less than that of the total Ca^{2+}
 332 contained in the membrane. Additional leaching tests in the presence of high salinity

333 (0.1 or 1 M NaCl) or a strong chelating agent (1, 10, or 100 mM
334 ethylenediaminetetraacetic acid (EDTA)) show the membrane was able to maintain its
335 stability (Table S6 in Supporting Information S5). Our results imply the formation of
336 highly stable Ca^{2+} -carboxyl complexes, which is likely caused by the binding of Ca^{2+}
337 to multiple $-\text{COO}^-$ sites within the membrane. According to Stumm and Morgen⁵⁵,
338 metal binding to adjacent ligands can increase the stability constant by orders of
339 magnitude.

340

341 **IMPLICATIONS**

342 We developed a simple and green modification method for preparation of antifouling
343 TFC membranes. For the first time, we show compelling evidence of significantly
344 improved antifouling performance by forming Ca^{2+} -carboxyl intra-bridging, which
345 sequestered these venerable groups to mitigate fouling. The Ca^{2+} -modified membrane
346 also exhibits enhanced surface hydrophilicity and water permeability without scarifying
347 its rejection. An added advantage of the intra-bridging is that the addition of CaCl_2
348 during interfacial polymerization can be easily realized within existing membrane
349 production lines. Thus, the intra-bridging approach provides a simple and cost-effective
350 way to overcome the longstanding permeability-rejection-antifouling tradeoff^{45, 56, 57}

351 In the current study, the Ca^{2+} -modified membranes show high stability even in the
352 presence of a strong chelating agent EDTA, which implies that Ca^{2+} is tightly bound in
353 the polyamide matrix by forming complex with multiple carboxyl groups. This type of

354 carboxyl- Ca^{2+} -carboxyl bonding is critical to make both calcium and carboxyl groups
355 unavailable to fouling. In contrast, pre-soaking a non-modified TFC membrane in
356 CaCl_2 is not effective for fouling mitigation (Supporting Information 4.3). In the latter
357 case, the absorbed Ca^{2+} , if not bound to multiple carboxyl groups in the polyamide, can
358 still form membrane- Ca^{2+} -foulant inter-bridges. Future studies may explore alternative
359 methods (such as the use of other divalent or multivalent metal ions) to further improve
360 the carboxyl sequestration efficiency.

361

362 **ASSOCIATED CONTENT**

363 The Supporting Information is available free of charge on the ACS Publications website:

364 S1. The casting solution compositions of Ca^{2+} -modified membranes; S2.
365 Characterization of the Ca^{2+} -modified membranes; S3. Forward osmosis (FO)
366 results; S4. Additional evidences of the sequestration effect for the Ca^{2+} -modified
367 membranes; S5. Membrane stability.

368

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374 **Notes**

375 The authors declare no competing financial interest.

376

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REFERENCES

1. Elimelech, M.; Phillip, W. A. The future of seawater desalination: Energy, technology, and the environment. *Science* **2011**, 333 (6043), 712-717.
2. Misdan, N.; Lau, W. J.; Ismail, A. F. Seawater Reverse Osmosis (SWRO) desalination by thin-film composite membrane-Current development, challenges and future prospects. *Desalination* **2012**, 287, 228-237.
3. Tang, C. Y.; Yang, Z.; Guo, H.; Wen, J. J.; Nghiem, L. D.; Cornelissen, E. Potable Water Reuse through Advanced Membrane Technology. *Environ. Sci. Technol.* **2018**, 52 (18), 10215-10223.
4. Tang, C. Y.; Chong, T. H.; Fane, A. G. Colloidal interactions and fouling of NF and RO membranes: A review. *Adv. Colloid. Interfac.* **2011**, 164, 126-143.
5. Guo, W.; Ngo, H. H.; Li, J. A mini-review on membrane fouling. *Bioresour. Technol.* **2012**, 122, 27-34.
6. Li, X.; Cai, T.; Chung, T. S. Anti-Fouling behavior of hyperbranched polyglycerol-grafted poly(ether sulfone) hollow fiber membranes for osmotic power generation. *Environ. Sci. Technol.* **2014**, 48 (16), 9898-9907.
7. Mo, Y.; Tiraferri, A.; Yip, N. Y.; Adout, A.; Huang, X.; Elimelech, M. Improved antifouling properties of polyamide nanofiltration membranes by reducing the density of surface carboxyl groups. *Environ. Sci. Technol.* **2012**, 46 (24), 13253-13261.
8. Tiraferri, A.; Elimelech, M. Direct quantification of negatively charged functional groups on membrane surfaces. *J. Membr. Sci.* **2012**, 389, 499-508.
9. She, Q.; Wang, R.; Fane, A. G.; Tang, C. Y. Membrane fouling in osmotically driven membrane processes: A review. *J. Membr. Sci.* **2016**, 499, 201-233.
10. Zhang, R.; Liu, Y.; He, M.; Su, Y.; Zhao, X.; Elimelech, M.; Jiang, Z. Antifouling membranes for sustainable water purification: Strategies and mechanisms. *Chem. Soc. Rev.* **2016**, 45 (21), 5888-5924.
11. Liu, C.; Faria, A. F.; Ma, J.; Elimelech, M., Mitigation of Biofilm Development on Thin-Film Composite Membranes Functionalized with Zwitterionic Polymers and Silver Nanoparticles. *Environ. Sci. Technol.* **2017**, 51 (1), 182-191.
12. Qin, D.; Liu, Z.; Liu, Z.; Bai, H.; Sun, D. D. Superior Antifouling Capability of Hydrogel Forward Osmosis Membrane for Treating Wastewaters with High Concentration of Organic Foulants. *Environ. Sci. Technol.* **2018**, 52 (3), 1421-1428.
13. Mi, B.; Elimelech, M. Organic fouling of forward osmosis membranes: Fouling reversibility and cleaning without chemical reagents. *J. Membr. Sci.* **2010**, 348 (1-2), 337-345.
14. Tang, C. Y.; Kwon, Y.-N.; Leckie, J. O. Probing the nano- and micro-scales of reverse osmosis membranes—A comprehensive characterization of physiochemical properties of uncoated and coated membranes by XPS, TEM, ATR-FTIR, and streaming potential measurements. *J. Membr. Sci.* **2007**, 287 (1), 146-156.
15. Wang, Z.; Elimelech, M.; Lin, S. Environmental Applications of Interfacial Materials with Special Wettability. *Environ. Sci. Technol.* **2016**, 50 (5), 2132-2150.

- 423 16. Tang, C. Y.; Kwon, Y.-N.; Leckie, J. O. Effect of membrane chemistry and coating
424 layer on physiochemical properties of thin film composite polyamide RO and NF
425 membranes. I. FTIR and XPS characterization of polyamide and coating layer
426 chemistry. *Desalination* **2009**, *242* (1-3), 149-167.
- 427 17. Li, Q.; Elimelech, M. Organic fouling and chemical cleaning of nanofiltration
428 membranes: Measurements and mechanisms. *Environ. Sci. Technol.* **2004**, *38* (17),
429 4683-4693.
- 430 18. Xiang, Y.; Liu, Y.; Mi, B.; Leng, Y. Molecular dynamics simulations of polyamide
431 membrane, calcium alginate gel, and their interactions in aqueous solution. *Langmuir*
432 **2014**, *30* (30), 9098-9106.
- 433 19. Lu, X.; Romero-Vargas Castrillón, S.; Shaffer, D. L.; Ma, J.; Elimelech, M. In situ
434 surface chemical modification of thin-film composite forward osmosis membranes for
435 enhanced organic fouling resistance. *Environ. Sci. Technol.* **2013**, *47* (21), 12219-
436 12228.
- 437 20. Tang, C. Y.; Kwon, Y.-N.; Leckie, J. O. Fouling of reverse osmosis and
438 nanofiltration membranes by humic acid -- Effects of solution composition and
439 hydrodynamic conditions. *J. Membr. Sci.* **2007**, *290* (1-2), 86-94.
- 440 21. Parida, V.; Ng, H. Y. Forward osmosis organic fouling: Effects of organic loading,
441 calcium and membrane orientation. *Desalination* **2013**, *312*, 88-98.
- 442 22. Motsa, M. M.; Mamba, B. B.; D'Haese, A.; Hoek, E. M. V.; Verliefde, A. R. D.
443 Organic fouling in forward osmosis membranes: The role of feed solution chemistry
444 and membrane structural properties. *J. Membr. Sci.* **2014**, *460*, 99-109.
- 445 23. Lee, S.; Elimelech, M. Relating organic fouling of reverse osmosis membranes to
446 intermolecular adhesion forces. *Environ. Sci. Technol.* **2006**, *40* (3), 980-987.
- 447 24. Mi, B.; Elimelech, M. Chemical and physical aspects of organic fouling of forward
448 osmosis membranes. *J. Membr. Sci.* **2008**, *320* (1-2), 292-302.
- 449 25. Jin, X.; Huang, X.; Hoek, E. M. V. Role of specific ion interactions in seawater RO
450 membrane fouling by alginic acid. *Environ. Sci. Technol.* **2009**, *43* (10), 3580-3587.
- 451 26. Zirehpour, A.; Rahimpour, A.; Arabi Shamsabadi, A.; Sharifian, M. G.; Soroush,
452 M. Mitigation of Thin-Film Composite Membrane Biofouling via Immobilizing Nano-
453 Sized Biocidal Reservoirs in the Membrane Active Layer. *Environ. Sci. Technol.* **2017**,
454 *51* (10), 5511-5522.
- 455 27. Perreault, F.; Jaramillo, H.; Xie, M.; Ude, M.; Nghiem, L. D.; Elimelech, M.
456 Biofouling Mitigation in Forward Osmosis Using Graphene Oxide Functionalized
457 Thin-Film Composite Membranes. *Environ. Sci. Technol.* **2016**, *50* (11), 5840-5848.
- 458 28. Baoxia, M. I.; Elimelech, M. Gypsum scaling and cleaning in forward osmosis:
459 Measurements and mechanisms. *Environ. Sci. Technol.* **2010**, *44* (6), 2022-2028.
- 460 29. Xie, M.; Gray, S. R. Gypsum scaling in forward osmosis: Role of membrane
461 surface chemistry. *J. Membr. Sci.* **2016**, *513*, 250-259.
- 462 30. Liu, C.; Lee, J.; Ma, J.; Elimelech, M. Antifouling Thin-Film Composite
463 Membranes by Controlled Architecture of Zwitterionic Polymer Brush Layer. *Environ.*
464 *Sci. Technol.* **2017**, *51* (4), 2161-2169.

- 465 31. Zhang, X.; Tian, J.; Gao, S.; Zhang, Z.; Cui, F.; Tang, C. Y. In situ surface
466 modification of thin film composite forward osmosis membranes with sulfonated
467 poly(arylene ether sulfone) for anti-fouling in emulsified oil/water separation. *J. Membr.*
468 *Sci.* **2017**, 527, 26-34.
- 469 32. Xu, W.; Ge, Q. Novel functionalized forward osmosis (FO) membranes for FO
470 desalination: Improved process performance and fouling resistance. *J. Membr. Sci.*
471 **2018**, 555, 507-516.
- 472 33. Zhang, X.; Tian, J.; Ren, Z.; Shi, W.; Zhang, Z.; Xu, Y.; Gao, S.; Cui, F. High
473 performance thin-film composite (TFC) forward osmosis (FO) membrane fabricated on
474 novel hydrophilic disulfonated poly(arylene ether sulfone) multiblock
475 copolymer/polysulfone substrate. *J. Membr. Sci.* **2016**, 520, 529-539.
- 476 34. Nghiem, L. D.; Schäfer, A. I.; Elimelech, M. Removal of Natural Hormones by
477 Nanofiltration Membranes: Measurement, Modeling, and Mechanisms. *Environ. Sci.*
478 *Technol.* **2004**, 38 (6), 1888-1896.
- 479 35. Yang, L.; She, Q.; Wan, M. P.; Wang, R.; Chang, V. W. C.; Tang, C. Y. Removal
480 of haloacetic acids from swimming pool water by reverse osmosis and nanofiltration.
481 *Water Res.* **2017**, 116, 116-125.
- 482 36. Guo, H.; Deng, Y.; Yao, Z.; Yang, Z.; Wang, J.; Lin, C.; Zhang, T.; Zhu, B.; Tang,
483 C. Y. A highly selective surface coating for enhanced membrane rejection of endocrine
484 disrupting compounds: Mechanistic insights and implications. *Water Res.* **2017**, 121,
485 197-203.
- 486 37. Zhao, H.; Qiu, S.; Wu, L.; Zhang, L.; Chen, H.; Gao, C. Improving the performance
487 of polyamide reverse osmosis membrane by incorporation of modified multi-walled
488 carbon nanotubes. *J. Membr. Sci.* **2014**, 450, 249-256.
- 489 38. Valladares Linares, R.; Yangali-Quintanilla, V.; Li, Z.; Amy, G. NOM and TEP
490 fouling of a forward osmosis (FO) membrane: Foulant identification and cleaning. *J.*
491 *Membr. Sci.* **2012**, 421-422, 217-224.
- 492 39. Liu, Y.; Mi, B. Combined fouling of forward osmosis membranes: Synergistic
493 foulant interaction and direct observation of fouling layer formation. *J. Membr. Sci.*
494 **2012**, 407-408, 136-144.
- 495 40. Ye, G.; Lee, J.; Perreault, F.; Elimelech, M. Controlled Architecture of Dual-
496 Functional Block Copolymer Brushes on Thin-Film Composite Membranes for
497 Integrated "defending" and "attacking" Strategies against Biofouling. *ACS Appl. Mater.*
498 *Inter.* **2015**, 7 (41), 23069-23079.
- 499 41. Ma, X.-H.; Yao, Z.-K.; Yang, Z.; Guo, H.; Xu, Z.-L.; Tang, C. Y.; Elimelech, M.
500 Nanofoaming of polyamide desalination membranes to tune permeability and
501 selectivity. *Environ. Sci. Technol. Lett.* **2018**, 5 (2), 123-130.
- 502 42. Ma, X.; Yang, Z.; Yao, Z.; Guo, H.; Xu, Z.; Tang, C. Y. Tuning roughness features
503 of thin film composite polyamide membranes for simultaneously enhanced
504 permeability, selectivity and anti-fouling performance. *J. Colloid. Interf. Sci.* **2019**, 540,
505 382-388.

- 506 43. Tang, C. Y.; Kwon, Y. N.; Leckie, J. O. The role of foulant-foulant electrostatic
507 interaction on limiting flux for RO and NF membranes during humic acid fouling-
508 Theoretical basis, experimental evidence, and AFM interaction force measurement. *J.*
509 *Membr. Sci.* **2009**, 326 (2), 526-532.
- 510 44. Wan, C. F.; Yang, T.; Gai, W.; Lee, Y. D.; Chung, T. S. Thin-film composite
511 hollow fiber membrane with inorganic salt additives for high mechanical strength and
512 high power density for pressure-retarded osmosis. *J. Membr. Sci.* **2018**, 555, 388-397.
- 513 45. Tang, C. Y.; Kwon, Y.-N.; Leckie, J. O. Effect of membrane chemistry and coating
514 layer on physiochemical properties of thin film composite polyamide RO and NF
515 membranes II. Membrane physiochemical properties and their dependence on
516 polyamide and coating layers. *Desalination* **2009**, 242 (1-3), 168-182.
- 517 46. Fane, A. G.; Tang, C. Y.; Wang, R. Membrane Technology for Water:
518 Microfiltration, Ultrafiltration, Nanofiltration, and Reverse Osmosis. *Treatise on Water*
519 *Science*, Wilderer, P., Ed. Academic Press: Oxford, **2011**; pp 301-335.
- 520 47. Wei, J.; Qiu, C.; Tang, C. Y.; Wang, R.; Fane, A. G. Synthesis and characterization
521 of flat-sheet thin film composite forward osmosis membranes. *J. Membr. Sci.* **2011**,
522 372, 292-302.
- 523 48. Yip, N. Y.; Tiraferri, A.; Phillip, W. A.; Schiffman, J. D.; Elimelech, M. High
524 performance thin-film composite forward osmosis membrane. *Environ. Sci. Technol.*
525 **2010**, 44 (10), 3812-3818.
- 526 49. Sukitpaneenit, P.; Chung, T. S. High performance thin-film composite forward
527 osmosis hollow fiber membranes with macrovoid-free and highly porous structure for
528 sustainable water production. *Environ. Sci. Technol.* **2012**, 46 (13), 7358-7365.
- 529 50. Liu, Q.; Qiu, G.; Zhou, Z.; Li, J.; Amy, G. L.; Xie, J.; Lee, J. Y. An Effective
530 Design of Electrically Conducting Thin-Film Composite (TFC) Membranes for Bio and
531 Organic Fouling Control in Forward Osmosis (FO). *Environ. Sci. Technol.* **2016**, 50
532 (19), 10596-10605.
- 533 51. Lu, X.; Arias Chavez, L. H.; Romero-Vargas Castrillón, S.; Ma, J.; Elimelech,
534 M. Influence of active layer and support layer surface structures on organic fouling
535 propensity of thin-film composite forward osmosis membranes. *Environ. Sci. Technol.*
536 **2015**, 49 (3), 1436-1444.
- 537 52. Xie, M.; Bar-Zeev, E.; Hashmi, S. M.; Nghiem, L. D.; Elimelech, M. Role of
538 Reverse Divalent Cation Diffusion in Forward Osmosis Biofouling. *Environ. Sci.*
539 *Technol.* **2015**, 49 (22), 13222-13229.
- 540 53. Du, Y. J.; Wei, M. L.; Reddy, K. R.; Liu, Z. P.; Jin, F. Effect of acid rain pH on
541 leaching behavior of cement stabilized lead-contaminated soil. *J. Hazard. Mater.* **2014**,
542 271, 131-140.
- 543 54. Song, F.; Gu, L.; Zhu, N.; Yuan, H. Leaching behavior of heavy metals from
544 sewage sludge solidified by cement-based binders. *Chemosphere* **2013**, 92 (4), 344-350.
- 545 55. Stumm, W.; Morgan, J. J. *Aquatic Chemistry, Chemical Equilibria and Rates in*
546 *Natural Waters*. 3rd Edition ed.; John Wiley & Sons: New York, 1996.

- 547 56. Meng, J.; Cao, Z.; Ni, L.; Zhang, Y.; Wang, X.; Zhang, X.; Liu, E. A novel salt-
548 responsive TFC RO membrane having superior antifouling and easy-cleaning
549 properties. *J. Membr. Sci.* **2014**, *461*, 123-129.
- 550 57. Werber, J. R.; Deshmukh, A.; Elimelech, M. The Critical Need for Increased
551 Selectivity, Not Increased Water Permeability, for Desalination Membranes. *Environ.*
552 *Sci. Technol. Lett.* **2016**, *3* (4), 112-120.

553